

1 **Climate and phenology drive coherent regional**
2 **interannual variability of carbon dioxide flux**
3 **in a heterogeneous landscape**

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18 **Abstract**

19 The climate sensitivity of plant seasonal life cycles, or phenology, may impart significant carbon
20 cycle feedbacks on climatic change. Analysis of interannual ecosystem carbon exchange
21 provides one way to assess this climate sensitivity. Multi-year eddy covariance carbon dioxide
22 flux observations from five ecosystems in the Upper Great Lakes USA, located 400 km of each
23 other and exhibiting coherent interannual variability, were used to parameterize a simple
24 ecosystem model. The model, when properly constrained with an interannual sensitive cost
25 function, was able to explain a significant proportion interannual variation of carbon fluxes in all
26 ecosystems except the old-growth forest. The results reveal that spring or autumn climate
27 thresholds impact annual carbon uptake, though the magnitude and strength varied by site. When
28 the model was forced to maintain the same climate-phenology relationship across the five sites,
29 most of the interannual variability could still be explained at most sites except the old-growth
30 forest and the forest furthest in distance from the others. These results suggest that coarse spatial
31 resolution carbon-climate models could likely specify general climate-phenological relationships
32 at grid scales on order of 100 km without appreciably sacrificing ability to model interannual
33 carbon cycling.

34 **Index terms:** Biogeochemical cycles, processes, and modeling (0414); Carbon cycling (0428);
35 Biosphere/atmosphere interactions (0426)

36 **Keywords:** Phenology; eddy covariance; ChEAS

37

37 **1. Introduction**

38 One key to understanding the impacts of the terrestrial carbon cycle on future climate change is
39 better diagnosis of climatic controls on interannual variability (IAV) of land-atmosphere carbon
40 dioxide net ecosystem exchange (NEE). This is especially the case in temperate and boreal
41 forests where IAV is large (Yuan *et al.*, 2009) and strongly linked to climate variability (e.g.,
42 Barr *et al.*, 2006; Chen *et al.*, 1999; Goulden *et al.*, 1996; Hollinger *et al.*, 2004; Sierra *et al.*,
43 2009). Large IAV could, for example, lead to reduced long-term carbon accumulation in some
44 forests, due to the impact of disturbance-driven respiration pulses (Sierra *et al.*, 2009). Despite its
45 importance, observed IAV of NEE in these biomes is difficult to capture in ecosystem models,
46 which are better tuned to capture diurnal, seasonal, and successional patterns (Ricciuto *et al.*,
47 2008; Urbanski *et al.*, 2007; Stoy *et al.*, 2009).

48 Currently, we lack a strong physical basis for many of the complex interactions that exist in
49 terrestrial systems at this timescale (Bonan, 2008; Stoy *et al.*, 2009). For example, while annual
50 NEE typically declines with latitude in temperate regions, relative IAV increases in deciduous
51 forests and declines in evergreen forests, a result that is difficult to explain (Yuan *et al.*, 2009). In
52 several grasslands, it was noted that sensitivity of plant productivity to climate drivers varied
53 year-to-year (Polley *et al.*, 2010). This result is similar to findings of Richardson *et al.* (2007)
54 who argued that biotic, not climate, variability was the primary cause of decadal flux variability
55 in a spruce forest.

56 Further, we lack understanding about what scales do we expect spatial coherence in interannual
57 variation, which may influence optimal scales of specifying climate-ecosystem relationships in
58 land-atmosphere models. One promising avenue of research for better modeling of regional

59 ecosystem model IAV is improved simulation of climate sensitivity in plant phenological life
60 cycles (Peñuelas *et al.*, 2009). Phenology links climate anomalies, especially in the shoulders of
61 the plant growing season, to plant biogeochemistry (Morissette *et al.*, 2009; Piao *et al.*, 2008).
62 Recent climatic warming leading to advances in spring flowering and leaf timing has been noted
63 in many parts of the globe (Linderholm, 2006) especially in Europe (Menzel *et al.*, 1999; Stöckli
64 and Vidale, 2004) and North America (Myneni *et al.*, 1997; White *et al.*, 2009), across diverse
65 ecosystems including temperate forests (e.g., Richardson *et al.*, 2006; Vitasse *et al.*, 2009) and
66 Mediterranean shrublands (e.g., Gordo and Sanz, 2010). Climate records indicate shifts in both
67 phase and amplitude of the annual temperature (Stine *et al.*, 2009), suggesting that links between
68 phenology and climate will likely have significant impacts on ecosystem productivity with
69 ensuing anthropogenic climate change. The impact of warm springs and longer growing season
70 lengths on carbon uptake has been well-noted at several sites (Barr *et al.*, 2006; Chen *et al.*,
71 1999; Churkina *et al.*, 2005; Goulden *et al.*, 1996; Hollinger *et al.*, 2004), but ecosystem models
72 that can capture this impact on IAV are elusive (Baldocchi *et al.*, 2005).

73 One way to advance our understanding is to develop and test ecosystem models constrained by
74 multi-year observations that connect phenology, carbon cycling, and climate at multiple sites.
75 Long-term eddy covariance flux towers, which directly observe NEE of ecosystems over
76 multiple years, are particularly well suited for testing how well models of phenology capture
77 carbon cycle IAV (Richardson *et al.*, 2009), though only a few studies have used multiple flux
78 towers (Baldocchi *et al.*, 2005; Churkina *et al.*, 2005) and none focused on multiple towers in
79 one region. Coherent IAV has been observed across a set of flux towers in a similar climate and
80 biome (Desai *et al.*, 2008), but there has been limited success in modeling this IAV. Here, I ask

81 to what extent can the observed IAV be explained by a simple climate sensitive model of plant
82 phenology and what does it imply for improving ecosystem IAV modeling?

83 To investigate this question, a simple ecosystem model was developed and parameterized using
84 Bayesian techniques against multi-year flux tower data observed in five ecosystems. Since
85 ecosystem model interannual variability can be strongly sensitive to how parameter estimation is
86 designed, an alternative IAV-sensitive formulation of the cost function was also investigated.
87 Finally, to investigate controls on synchronous forcing, model parameterization was further
88 modified to force spatial convergence on phenological parameters. Findings from these
89 investigations are used to discuss implications for environmental controls and spatial coherence
90 of regional IAV.

91 **2. Data and methods**

92 ***2.1 Site and data description***

93 Eddy covariance flux towers in the temperate-boreal transition region of the Upper Great Lakes
94 were analyzed in this study (Table 1). The sites, which included three forests, one shrub wetland,
95 and one tall tower regional mixed forest-wetland footprint, were located within 400 km of each
96 other (Fig. 1), each have at least five years of flux and meteorological data, and have been
97 previously analyzed and described in an upscaling study (Desai *et al.*, 2008). Four of the sites
98 (US-WCr, US-Syv, US-Los, and US-PFa) are within 150 km of each other in north central
99 Wisconsin/upper Michigan, while the fifth site, US-UMB, is in northern Lower Michigan. Of the
100 forest sites, two are mature age class (US-WCr and US-UMB), and one is old-growth (US-Syv).
101 The wetland site is a short-stature shrub alder-willow fen. The tall tower is a 447-m radio tower

102 with flux measurements at three heights. For the purposes of this comparison, the fluxes from the
103 three levels of the tall tower are combined in an optimal selection strategy, as described by Davis
104 *et al.* (2003), to produce a single “regional” NEE.

105 Meteorological data at each site were acquired and gap-filled using a combination of nearest
106 neighbor and moving-window ensemble diurnal average techniques (Desai *et al.*, 2008). Eddy
107 covariance and storage fluxes of CO₂ were used to compute NEE at each site. Standard flux
108 computation methods at each site were relatively similar (Desai *et al.*, 2008) and fluxes
109 computed by these codes have compared favorably to the Ameriflux “gold” standard, a network-
110 wide blind data processing protocol. Common techniques were used in screening for low
111 turbulence conditions and gap filling of data gaps that occur due to low turbulence or instrument
112 failure (Desai *et al.*, 2005). The gap-filling technique compared well with other standard
113 techniques used by the flux tower community (Moffat *et al.*, 2007). All fluxes were computed at
114 the half-hourly scale, except for US-PFa, which used an hourly scale due to the taller height. For
115 assimilation into the model, all flux and meteorological data were averaged across day and night
116 periods, similar to the method of Sacks *et al.* (2006). Using half-daily summed fluxes reduces
117 impact of random turbulent flux error on data assimilation, but retains the nocturnal respiration
118 signal. Summed half-daily flux integrals whose hours were more than 25% gap-filled were
119 discarded for data assimilation to minimize artifacts arising from model-model comparison.

120 **2.2 Model description**

121 A simple ecosystem model, the Interannual Flux Tower Upscaling Experiment (IFUSE), was
122 parameterized against all site data. The model consisted of 17 total parameters (Table 2), of
123 which 3 were fixed for each site, 10 were optimized at each site, and four phenology parameters

124 were either optimized at each site separately (asynchronous mode) or jointly for all sites
125 (synchronous mode). The model was run at a half daily (day/night) adaptive length time step,
126 which has been shown to be well suited for parameter optimization against flux tower NEE
127 (Sacks *et al.*, 2006).

128 At each time step, the model applied environmental forcing of canopy air temperature (T_a), 5 cm
129 soil temperature (T_s), photosynthetic active radiation (PAR), and vapor pressure deficit (VPD) to
130 estimate gross primary production (GPP), ecosystem respiration (ER), and NEE in gC m^{-2}
131 timestep^{-1} and leaf area index (LAI) in $\text{m}^2 \text{m}^{-2}$. GPP was estimated using a five-parameter light,
132 temperature, and VPD limited modified light use efficiency equation:

$$133 \quad GPP = LUE \cdot \left(1 - e^{-k \cdot LAI}\right) PAR \left(\frac{T_a - T_{\min}}{T_{\text{opt}} - T_{\min}}\right) \left(\frac{VPD_{\max} - VPD}{VPD_{\max} - VPD_{\min}}\right) \quad (1)$$

134 where LUE, k , T_{\min} , T_{opt} , VPD_{\max} , and VPD_{\min} are model parameters as described in Table 2.

135 To calculate LAI, leaf phenology of emergence and senescence was modeled with a two-
136 parameter sigmoidal relationship. The phenology model used here consisted of the well-
137 established accumulated growing degree days base 10 C (GDD) approach for mid-point of leaf
138 emergence and a 5 cm soil temperature threshold for mid-point of leaf senescence, models which
139 have been shown to explain much of the variation in canopy development for northern forests
140 (Baldocchi *et al.*, 2005; Richardson *et al.*, 2006). Canopy fraction with evergreen vegetation was
141 simulated by preventing LAI to decline beyond a minimum threshold (LAI_{\min}), leading to:

$$\begin{cases}
LAI = LAI_{\min} + [(LAI_{\max} - LAI_{\min}) L_{spring} L_{fall}] \\
L_{spring} = \frac{1}{e^{-\alpha \left(\frac{DOY - L_{on}}{2} \right)}}; L_{fall} = 1 - \frac{1}{e^{-\beta \left(\frac{DOY - L_{off}}{2} \right)}} \\
L_x = \frac{L_x - \min(L_x)}{\max(L_x) - \min(L_x)}; x = spring / fall \\
L_{on} = DOY|_{GDD > GDD_{Thresh}}; L_{off} = DOY|_{T_s < T_{Thresh} \cup L_{spring} > 0.99}
\end{cases} \quad (2)$$

143 where LAI_{\min} , LAI_{\max} , α , β , GDD_{thresh} , and T_{thresh} are model parameters (Table 2). L_{spring} and L_{fall}
144 describe the variation of LAI around the leaf on (L_{on}) or leaf off (L_{off}) day of year (DOY). Both
145 of these functions were further normalized to vary between 0 and 1 and then multiplied together,
146 thus allowing LAI to vary between LAI_{\min} and LAI_{\max} . Consequently, interannual variability in
147 LAI is quite muted in this model, which allowed this model to focus on the role of growing
148 season length (G_{SL}) on IAV of NEE.

149 ER was estimated with five parameters that control respiration rates in three soil pools sensitive
150 to T_s , T_a , and GPP:

$$ER = \left[r_s \cdot e^{b_1(T_s - 15)} \right] + \left[r_v \cdot e^{b_2(T_a - 15)} \right]_{GPP > 0} + \left[b_3 \cdot GPP_{DOY-1} \right] \quad (3)$$

152 where r_s , r_v , b_1 , b_2 , and b_3 are model parameters (Table 2). The first term represents the combined
153 effects soil heterotrophic and plant maintenance respiration. The second term, which is only
154 present when GPP is positive, represents plant growth respiration (or alternatively, the change in
155 respiration sensitivity in growing vs. dormant seasons), while the final term, is a fraction of the
156 previous day GPP, representing autotrophic respiration of newly assimilated carbohydrate and
157 allows for GPP lag effects. These formulations were chosen to represent the dynamics that could

158 likely be resolved from NEE measurements, as opposed to a more mechanistic, but also more
159 parameter intensive and pool sensitive model of respiration.

160 NEE was computed as the residual between ER and GPP. The model was designed to be
161 purposefully simple so as to capture the key diurnal and season dynamics typically seen in NEE
162 observations while limiting the number of parameters. By avoiding specification of soil and
163 biomass pools outside of leaves, the model removed one of the largest sources of uncertainty and
164 equifinality in estimating model parameters from flux data (Luo *et al.*, 2009). This structure
165 implied that soil pools were assumed to be steady state relative to the fluxes, which is likely a
166 reasonable assumption for mature, established secondary succession ecosystems. Since the goal
167 was simulation of daily to interannual NEE at timescales less than a decade (i.e., <10% of a
168 temperate hardwood forest successional cycle), this assumption implied that short-term
169 variations in NEE were driven entirely by the response of ER and GPP to climate. Given the
170 mesic climate of the region, influences of precipitation and moisture variability were assumed to
171 be minimal. The impact of these assumptions on interpretation of results is provided in the
172 discussion.

173 ***2.3 Model parameterization***

174 Free parameters of the model (phenology, photosynthesis and respiration parameters in Table 2)
175 were estimated using a Markov Chain Monte Carlo (MCMC) estimator (Braswell *et al.*, 2005)
176 with the Metropolis-Hasting algorithm (Metropolis and Ulam, 1949). In this approach, free
177 parameters were randomly perturbed across a range of reasonable prior values, assuming a
178 uniform distribution (Table 2). New parameter sets were “accepted” when a cost function
179 indicated better fit of model to data, and occasionally when not, so as to avoid local minima.

180 Multiple chains (six, in this study) were built from random locations in parameter space and
181 iterated until a convergence criterion is reached, usually within 50,000 iterations. Iteration sizes
182 were chosen to be arbitrarily large, likely oversampling the parameter space. The best chain was
183 then propagated forward another 70,000 iterations, and a subset of the final 80,000 iterations
184 were saved as “accepted” parameter sets based on the acceptance criterion. Best model output
185 and variance were computed from the model output of these accepted parameter sets. More
186 details of the general approach are provided in Braswell *et al.* (2005).

187 The first five years of half-daily NEE observations for each site were used in the MCMC cost
188 function to minimize model-data mismatch. The cost function can be written as:

$$189 \quad L_D = \prod_{i=1}^n \frac{1}{\sqrt{2\pi\sigma}} e^{-\frac{(x_i - \mu_i)^2}{2\sigma^2}} \quad (4)$$

190 where L_D is the likelihood to be minimized, x_i is observed half-daily NEE, μ_i is model NEE, and
191 σ^2 is data error with respect to model structure, which was computed as the mean sum of square
192 deviations between x_i and μ_i (Sacks *et al.*, 2006). To improve numerical stability, this equation
193 was computed as the log likelihood, allowing the product function to be written as a sum.
194 Additionally, to test whether the cost function biases how well the model identifies parameters
195 responsible for determining carbon flux IAV, Eq. (4) was further modified to account for both
196 fast (half-daily) and slow (annual) variations in NEE:

$$\left\{ \begin{array}{l}
 L_y = \prod_{i=1}^n \frac{1}{\sqrt{2\pi}\sigma} e^{-\frac{(x_i^m - \mu_i^m)^2}{2\sigma^2}} \\
 x_i^m = \sum_{j=i-DOY}^i x_j; \mu_i^m = \sum_{j=i-DOY}^i \mu_j \quad (5) \\
 L = L_D L_y
 \end{array} \right.$$

197 where L_y is a likelihood for annual NEE, and x_i^m is observed cumulative NEE from the start of the
 198 year to point i , and μ_i^m is the equivalent for the model. The new cost function is then product of
 199 the two likelihoods. Thus the model trades fit at the daily scale for fits at the annual scale, with
 200 the assumption that accepted parameters sets would be Pareto optimal for both. Cumulative NEE
 201 was used instead of annual NEE to allow for both the daily and annual to have roughly the same
 202 weight and also to prevent the model from fitting annual NEE with a poor seasonal pattern. The
 203 expectation is that this cost function may improve model reproduction of seasonal NEE and IAV
 204 without significant loss in explaining short-term variation. While L_d was weighted to account for
 205 the influence of gap-filled data by removing half-daily NEE sums with more than 25% gap-
 206 filled, L_y included all NEE to create well-formed NEE integrals. In most cases, this did not add
 207 significantly more points or appear to bias results.
 208

209 In addition to the two cost functions above, an alternate optimization was performed where
 210 phenology parameters (Table 2) were forced to be the same for all five sites. In this “synchrony”
 211 setup, the parameter optimization was run in tandem at all five sites, such that the four phenology
 212 parameters were optimized to be the same at all five sites, while the other ten parameters were
 213 allowed to vary by site. Computationally, this was simply performed by concatenating arrays of

214 flux and forcing data for all sites, with 5 sets of 10 independent parameters (50), and 4 co-
215 dependent parameters. Thus a single MCMC run estimated 54 parameters, instead of 14 at a time
216 for each site. To compensate for the larger number of parameters, the number of chains was
217 increased to 20 and total iterations to 432,000 (with 144,000 spin-up iterations). Since phenology
218 parameters were jointly modified at all five sites, the optimized phenology parameters were the
219 same at all sites in the synchrony optimization.

220 ***2.4 Experimental design***

221 Given the two cost functions, and two forms of parameter optimization (asynchronous and
222 synchronous), a total of three out of four experiments or model modes were analyzed here. These
223 include asynchronous optimization with the Eq. (4) half-daily cost function (“A_H”), the
224 asynchronous optimization with the Eq. (5) interannual cost function (“A_I”), and finally a
225 synchronous optimization with the Eq. (5) cost function (“S”). Results from the fourth,
226 synchronous optimization with the daily cost function, were performed but not discussed here
227 since the results are nearly identical to the A_H experiment. The three experiments provide
228 information about how well IAV can be parameterized and simulated at five sites by a simple
229 ecosystem model (experiments A_H and A_I) as well as test how important synchronous
230 phenological forcing drives coherent IAV (experiment S).

231 **3. Results**

232 ***3.1 Seasonal patterns***

233 Mean annual NEE at the five sites ranged from a large carbon sink (negative) to a moderate
234 source (positive) of CO₂ to the atmosphere (Table 1). Mature forests such as US-WCr and US-

235 UMB were the largest sinks followed by the shrub wetland (US-Los). In contrast, the old-growth
236 forest (US-Syv) was near neutral (with large fluctuations between source and sink in any one
237 year), and the mixed regional very tall tower (US-PFa) was the largest source. While the regional
238 CO₂ source observed by the tower has been a continuing puzzle (see Davis *et al.*, 2003 and
239 Ricciuto *et al.*, 2008 for further discussion and Desai *et al.*, 2010 for a regional perspective), the
240 other towers have NEE in line with expectations for vegetation type and latitude (Yuan *et al.*,
241 2009). Uncertainty arising from random error, gap-filling and low-turbulence filtering was
242 generally small and similar at all sites.

243 Despite large variations in annual NEE, mean seasonal patterns at all five sites were quite similar
244 (Fig. 2, black line). Generally, sites started absorbing carbon in late May/early June, and crossed
245 zero in cumulative NEE by mid-June, and turned back into carbon sources by late August or
246 early September, except for US-PFa which turned into a carbon source much earlier in August.
247 While the dates were similar for each site, small variations in those dates led to large differences
248 in the length of carbon uptake period, with the largest at the mature forests and wetlands, and
249 shortest in the old-growth forest and regional site. This finding provides one basis that growing
250 season timing and length were a strong controlling factor of annual carbon flux, and hence IAV.

251 When compared to seasonal flux tower NEE, the IFUSE model (averaged across the same years
252 as observations) generally replicated this pattern regardless of optimization mode (Fig. 2, dotted
253 and gray lines). In all cases, the A_H experiment best replicated half-daily variations in NEE,
254 explaining 83%-93% of variability (Table 3), followed in most cases by A_I, except at US-UMB,
255 where the S model outperformed A_I. Both A_H and A_I performed worst at US-PFa, perhaps
256 because of the mixed footprint, while the S model performed worst at US-Syv. All modes
257 underestimated the strength of growing season uptake at US-PFa. Differences of model

258 performance due to change in cost function are discussed in the next section. Visually (Fig. 2), it
259 is apparent that the A_I model better captured seasonal variation than A_H at most sites, especially
260 at US-Syv, but for magnitude of annual NEE, all modes did well with respect to uncertainty in
261 observations, except for A_H at US-Syv.

262 **3.2. Interannual variability**

263 IAV at all sites was present and ranged from modest in the case of the wetland, to large in the
264 case of the mature and old-growth forests (Table 1). When mean annual NEE is subtracted from
265 observations and the anomalies are then normalized by standard deviation, consistent patterns
266 emerge among the sites (Fig. 3). At most sites, IAV rarely exceeded $1-\sigma$, with the exception of
267 2001, coincident with a large regional forest tent caterpillar outbreak (Cook *et al.*, 2008). Year-
268 to-year fluctuations were quite common, and hints of a longer decadal scale variability were
269 evident. The spatial coherence of these anomalies are discussed in the next section.

270 No relationship existed between NEE magnitude and $1-\sigma$ IAV (Table 3). In absolute terms, IAV
271 was smallest in the wetland, barely detectable within the uncertainty in mean NEE, and largest at
272 US-WCr, the site most impacted by the 2001 insect outbreak. However, in relative terms (ratio of
273 $1-\sigma$ IAV to NEE), the largest IAV was at US-Syv, where NEE was near neutral. In all cases,
274 IAV was a major fraction of NEE.

275 Despite all modes being able to capture most of the seasonal pattern of cumulative NEE at all
276 sites, simulation of IAV anomalies was poorly modeled at all sites by A_H (Fig. 4a), but
277 significantly improved in A_I (Fig. 4b). This result highlights the importance of the dual-
278 likelihood cost function used by A_I . A_I IAV anomalies were significantly correlated ($p < 0.05$) to

279 observed IAV anomalies at US-WCr, US-UMB, and US-PFa, strongly correlated ($p < 0.1$) to US-
280 Los, and weakly correlated to US-Syv (Table 3). The tradeoff, of course, was loss of explanation
281 of variance at the half-daily scale ranging from 3-11% and an increase in mean absolute error of
282 half-daily NEE by 0.03 to 0.3 $\text{gC m}^{-2} \text{day}^{-1}$.

283 Further evidence that the fit of A_I to observations is not just an artifact of modeling comes when
284 A_I L_{on} and L_{off} dates were compared to similar data observed at US-WCr. The observed dates
285 were derived from calibrating a simple LAI model to the ratio of above and below canopy
286 downwelling PAR and observed LAI (Cook *et al.*, 2008). While correlations are modest, the
287 model generally captured the pattern of anomalies in L_{ON} and L_{OFF} , though it did appear to
288 underestimate the variability in L_{on} and overestimate the variability in L_{off} . The net effect,
289 however, was good performance at simulating variations in growing season length.

290 In parameter space, there are a number of differences between A_H (Table 4) and A_I (Table 5).
291 Among phenology parameters, the net effect was in most but not all cases to increase the values
292 of phenology slope parameters, α and β , to higher values (steeply curved), consequently
293 modifying the phenology climate thresholds. Covariances between these parameters and
294 photosynthesis parameters then led to changes in LUE and temperature regulation of
295 photosynthesis, since these parameters can also act like phenology parameters. Less clear were
296 the reasons behind large changes in respiration parameters. While both A_H and A_I essentially
297 optimized to a value of zero for b_3 (fraction of GPP respired), A_I has significantly smaller b_2 (less
298 temperature sensitivity for plant growth respiration), and some large differences for b_1 for the
299 mature forest sites, suggesting equifinality in model solutions for respiration and highlighting the
300 difficulty of estimating optimal respiration parameters from eddy covariance data.

301 **3.3 Synchronous forcing**

302 While the magnitude of NEE and IAV varied across sites and with time, coherent anomalies in
303 IAV were readily observed when anomalies were statistically standardized (Fig. 3). These results
304 indicate the strong role that regional climate variability had on IAV. Even after taking NEE
305 uncertainty into account, coherence was strong, with most sites showing positive anomalies in
306 2001 and 2004, and negative anomalies in 1999-2000, and 2002-2003. A bifurcation of trends
307 across sites occurred in 2005 and 2006. While the tent caterpillar outbreak was one source of the
308 anomaly in 2001, the outbreak was mostly felt at US-WCr in late spring, to a small extent in the
309 US-PFa footprint, and barely at the other two sites, suggesting that climate anomalies still
310 explained most of the 2001 anomaly.

311 The S model tested whether synchronous IAV can be explained by coupling of phenological
312 parameters across sites. In this synchronous mode, the model still captured much of the IAV at
313 US-WCr, US-Los, and US-PFa ($p < 0.05$), but lost ability to simulate IAV at US-UMB, and like
314 all modes, could not capture IAV at US-Syv (Table 3). Correlation coefficients for the significant
315 correlations were essentially unchanged compared to A_I . It is interesting to note that US-UMB is
316 the further site from the rest (Fig. 1).

317 The trade-off in half-daily NEE simulation compared to A_H was quite similar to the trade-off
318 found for A_I , though with a larger drop in correlation at US-Syv. Across all sites, A_I has a strong
319 correlation with observed IAV ($r^2 = 0.81$) (Fig. 4b), but the S mode is still strongly correlated
320 ($r^2 = 0.68$), and with 16 fewer parameters in aggregate compared to A_I . Photosynthesis and
321 respiration parameters in A_I (Table 5) and S (Table 6) were more similar than between those and
322 A_H (Table 4). Remarkably, the parameters in S appeared more in line with literature estimates

323 than those for A_I , especially T_{opt} . Phenology parameters in S are fixed for all five sites, and
324 appeared to fall roughly near the average of those parameters for each site in A_I .

325 **4. Discussion**

326 **4.1 Modeling of IAV**

327 The observed IAV is within the range (~ 50 - $100 \text{ gC m}^{-2} \text{ yr}^{-1}$) observed for deciduous broadleaf
328 forests in mid-latitudes (Yuan *et al.*, 2009), with mature northern hardwood forests exhibiting the
329 largest. The results here contribute to findings of interannual variations in seasonal temperature
330 as a dominant driving force of interannual variation in carbon flux at mid-latitudes (Sierra *et al.*,
331 2009; Yuan *et al.*, 2009). Given the observed coherent IAV across space and likelihood that
332 seasonal temperature fluctuations were important in the study region, then, it is not entirely
333 surprising that a model tuned to capture daily to seasonal climate sensitivity of carbon cycling
334 can adequately capture the observed IAV, especially given that seasonal climate forcing strongly
335 influenced modeled ecosystem growing season length and timing. Rather, what is surprising is
336 the level of care needed for proper parameterization and the importance of not just identifying
337 optimal model structure and parameters, but also optimal model cost functions. In this case, as in
338 many ecosystem models, the failure of a simple MCMC approach in tuning parameters for
339 simulating IAV relies partly on the large signal imparted by CO_2 flux diurnal variability (large
340 and negative in day, large and positive at night), which tends to mask the more subtle, but
341 perhaps more climatically relevant, interannual signal (Stoy *et al.*, 2009).

342 The simulation presented here, consequently, is one of few models that have been able to
343 successfully diagnose interannual variability of NEE within a relatively simple ecosystem model.

344 Other well-known, and arguably more sophisticated models have shown less ability to model
345 IAV in north temperate forested regions (Ricciuto *et al.*, 2008; Urbanski *et al.*, 2007). It is likely,
346 at least in the case of the northeastern forest studied by Urbanski *et al.* (2007), that successional
347 trajectory was an important factor in long-term variability at the site, possibly overwhelming a
348 climate variability signal. A major shortcoming of the approach used here is the inability to
349 estimate how important succession and disturbance is part of sub-decadal IAV. Investigations at
350 decadal or longer timescales would certainly need to incorporate these process. Further, the
351 steady-state spin-up assumption made by developing a model with no carbon pools would not be
352 valid, and other techniques should be implemented with a pool based model (e.g., Carvalhais *et*
353 *al.*, 2008).

354 Analysis also revealed large variation in IAV variance explained by this model within a small
355 region that appeared to vary as a function of vegetation type, highlighting the importance of
356 individual site characteristics in determining the extent to which interannual carbon cycling may
357 be more controlled by climate or internal biotic dynamics (Polley *et al.*, 2010). Strong internal
358 control of NEE variability appears to have played a part in the lack of model predictive ability at
359 the old-growth forest. Desai *et al.* (2005) also highlighted the greater sensitivity to moisture
360 stress that has been found at this forest compared to nearby mature forests.

361 One way to assess biotic control that has been demonstrated both by Polley *et al.* (2010) and
362 Richardson *et al.* (2007), is to compare model parameterization with fixed parameters over
363 multiple years against interannually varying parameters, the latter reflecting variability in biotic
364 controls on NEE. Polley *et al.* (2010) argued that biotic control of interannual variability was
365 significant in grasslands, and Richardson *et al.* (2007a) similarly argued that the majority (55%)
366 of interannual variations in a spruce forest in the northeast US was driven by biotic variation.

367 Though this study did not test a model with variable parameters, the findings here of strong
368 explanation by a simple model with fixed-in-time parameters suggests that, at least for mature
369 hardwood forests of boreal-temperate transition regions, climate sensitivity, especially of spring
370 and fall, drove interannual variation of NEE.

371 While model was able to simulate IAV at the wetland, it was more designed with forest
372 productivity and aerobic decomposition in mind, suggesting non-shrub or precipitation-fed
373 wetlands would not fare as well as the shrub fen studied here. Still, this particular wetland site
374 was not in steady state over the time period due to a significant ongoing decline in water table
375 (Sulman *et al.*, 2009). Sulman *et al.* (2009) showed that water table influences both respiration
376 and productivity this wetland, generally leading to little change in NEE over the time period
377 studied. Consequently, one could argue that this model may be getting the right answer for the
378 wrong reasons, given the relatively small interannual variability of NEE.

379 Differing model structures led the MCMC algorithm to select different optimal values for many
380 number of parameters. When the models were compared in parameter space, it was not always
381 immediately obvious how other parameter differences between the models improved the fit to
382 IAV. These parameter correlations require further examination and suggest that caution is
383 required when drawing inferences from model parameter optimization techniques without first
384 testing for optimal model structure. Additionally, a question of overfitting to IAV arises when
385 relying on datasets with only a few years of data and the modified cost function, which also
386 requires further examination.

387 **4.2 Synchronous phenological forcing**

388 Another major finding of the work here is the how well a synchronously forced phenology
389 explained much of the IAV in the region for forests. Though the details of plant phenology vary
390 strongly by species and microclimate experienced by individual plants, at the scale of stand-level
391 carbon cycling, results here suggest that carbon cycle responses can generally be estimated by
392 relatively simple accumulated climatic heating indices and regional soil temperature thresholds.
393 Within this framework, it is not surprising that the US-UMB site fared poorest under
394 synchronous forcing, as it is the farthest site both climatically and geographically from the other
395 sites. These results also hint at a possible way to better estimate the spatial coherence of
396 phenological forcing by utilizing sets of flux towers to geostatistically test the ability of models
397 to jointly simulate flux variability. Here, the findings suggest synchronous scales of at least 100
398 km, reflecting the distance among the tower sites outside of US-UMB. Also, the decline in
399 explanation of variance at the old growth forest further develops the case that this site has strong
400 internal control on interannual carbon cycling.

401 **4.3 Carbon cycling and growing season length**

402 The parameterized A_I model can be further examined to suggest mechanisms that connect
403 climate variability to flux variability, via the interaction of model parameters that impact
404 growing season length (Fig. 7). For sites where the A_I model successfully simulated IAV, the
405 mechanism of by which phenology impacted IAV was not consistent across all sites. Hardwood
406 forest sites (US-WCr and US-UMB) showed less carbon uptake (more positive NEE) with later
407 L_{ON} , while other sites had no significant relationship (Fig. 7a). For these two sites, the strength of
408 this L_{ON} relationship drove a negative relationship between growing season length (G_{SL}) and

409 NEE (longer G_{SL} = more uptake). This finding is consistent with previous single site studies that
410 have noted relationships between warmer springs and enhanced annual carbon uptake in a boreal
411 aspen forest (Barr *et al.*, 2006; Chen *et al.*, 1999), eastern deciduous forest (Goulden *et al.*,
412 1996), and a spruce-dominated eastern forest (Hollinger *et al.*, 2004; Richardson *et al.*, 2009).

413 The effect of autumn (L_{OFF}) is less clear, with one only forest (US-Syv) showing a significant
414 positive relationship (later L_{OFF} = less uptake) (Fig. 7b). The wetland site (US-Los) also had a
415 significant relationship, but the magnitude was very small. A recent paper noted that warmer
416 autumns led to less carbon uptake in boreal ecosystems, by increasing ER more than GPP (Piao
417 *et al.*, 2008). This effect is not strongly evident here in the temperate-boreal transition zone.

418 While strong consistent spring and autumn climate impacts on NEE were not apparent, the effect
419 of both of these effects on G_{SL} is significant and negative (longer growing season = more carbon
420 uptake) at all sites except the mixed regional site (US-PFa), consistent with earlier findings
421 across the flux tower network showing growing season length as a strong determinant of net
422 carbon uptake (Baldocchi *et al.*, 2001; Churkina *et al.*, 2005). The lack of strong correlation at
423 the US-PFa site may be related to complementary responses occurring across the mix of stand
424 types sampled by the tall tower and perhaps the influence of moisture on regional fluxes that is
425 not apparent at the stand-scale towers (Desai *et al.*, 2010).

426 With respect to moisture, there is some evidence to suggest that relationships between water and
427 carbon cycle are an important factor on IAV to consider. Hu *et al.* (2010) found that evergreen
428 montaine forest carbon uptake had an inverse relationship with growing season length, due to the
429 importance of snowmelt as a source of growing season plant available water. The findings here,
430 showing mostly the opposite case, do not suggest a strong control of snow water on IAV in the
431 study region. However, other studies in the region have shown that water table depth (Desai *et*

432 *al.*, 2010; Sulman *et al.*, 2009) and summer soil moisture (Ricciuto *et al.*, 2008) may also be
433 important factors in explaining IAV in the patchy forest-wetland landscape that characterizes the
434 region, and in similar forests of other regions (Hollinger *et al.*, 2004). Ricciuto *et al.* (2008)
435 noted that daytime and seasonal NEE at the regional tall tower (US-PFa) were correlated to soil
436 moisture, but correlations were weak at the annual scale. The models used here did not consider
437 these effects, which may explain some of the unexplained variability of IAV, especially at the
438 wetland and old-growth forest. Time lags are likely in relationships between moisture and carbon
439 (e.g., Desai *et al.*, 2010; Dunn *et al.*, 2007; Hu *et al.*, 2010), and model mechanisms to couple
440 these processes require further assessment.

441 **5. Conclusion**

442 Thirty-one site years of near continuous flux tower carbon exchange observations across a meso-
443 network of five established Ameriflux sites were used to identify a coherent signal of interannual
444 variability in net ecosystem exchange, a likely indicator of the role of regional climate variability
445 on ecosystem carbon cycling. A model parameterized with climate-sensitive phenology and a
446 minimal set of carbon cycle functions and parameters to explain daily variations in NEE could
447 successfully simulate much of this IAV, especially at the mature forest sites, but only when the
448 model cost function was correctly identified and applied.

449 Climate variability in this boreal-temperate transition region drove NEE variability in the model
450 primarily through the impact of growing season length on length of carbon uptake period. These
451 results also suggest that timescales over which climate impacts decomposition and respiration is
452 longer, which is not surprising given the longer residence time of carbon in decomposing pools
453 (soil) versus photosynthetic pools (leaves). The model also highlighted the role that climate

454 variability imparts on carbon flux spatial coherence, at least on length scales of 100 km, though
455 this question would be best further explored with a larger scale geostatistical study of carbon flux
456 spatial variation.

457 Old-growth forest and wetland annual carbon flux variability were less well simulated,
458 suggesting a strong role for internal biotic dynamics and moisture variability on carbon flux
459 variations at some sites. These dynamics may be an important aspect of regional carbon cycle
460 variability, especially as forests in the region age and long-term drought conditions persist. Other
461 noted causes of IAV that also require further consideration, especially at regional scales, include
462 the role of stochastic disturbance (Desai *et al.*, 2007), pest outbreaks (Cook *et al.*, 2008), and
463 internal organic matter decomposition dynamics (Ricciuto *et al.*, 2008).

464 The impact of climate variability on phenology and ultimately ecosystem biogeochemistry is a
465 first order climate-ecosystem interaction, and of likely importance on the predictability of future
466 carbon cycles as anthropogenic climatic changes are expected to be strongly felt in mid-
467 continental mid-latitude regions. Preliminary findings from long term flux tower observations
468 and careful ecosystem model parameterization in a boreal-temperate transition region suggest
469 that future climate change in the shoulder seasons is likely to affect the carbon balance of mixed
470 and deciduous broadleaf forests, perhaps more than climatic changes occurring in the central part
471 of the growing season. However, these findings are limited by lack of longer-term carbon cycle
472 and phenological observations. Additionally, the variety of findings among montaine, grassland,
473 temperate, boreal, and temperate-boreal transition regions highlight the need for continued
474 efforts to better parameterize climate sensitivity of phenology in ecosystem models.

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611

611 **Figure captions**

612 **Figure 1.** Map of north-central United States showing location of the five flux tower sites used
613 in this study.

614

615 **Figure 2.** Ensemble average cumulative daily NEE for observations, and the three model
616 experiments (A_H , A_I , and S). Averages were performed over the observed record noted in Table
617 1. The model generally captured the seasonal pattern of NEE at all sites, with worst performance
618 at US-Syv and US-PFa. Observed and model uncertainty is not shown to preserve clarity, but
619 generally fall within 10% of any observation.

620

621 **Figure 3.** Observed standardized interannual variability in NEE at the five study sites. Strong
622 coherence in variability in NEE was observed across the time period, even though absolute
623 magnitudes in NEE variability varied widely. Observational uncertainty in NEE is noted by the
624 horizontal bars.

625

626 **Figure 4.** Correlation of anomalies in observed and modeled annual NEE using a) the A_H cost
627 function parameters (Table 4) and b) the A_I cost function parameters (Table 5). Significant
628 improvement in simulation of interannual variability was found for all sites in the latter.

629

630 **Figure 5.** Comparison of observed light extinction profile derived leaf on (square) and leaf off
631 (triangle) date anomalies to IFUSE model for the US-WCr site, using the A_I parameters.
632 Generally, variability in both dates was modestly well simulated, though the slope of leaf off
633 appears too steep, while the leaf on dates mostly fall on the 1:1 line except for one outlier.

634

635 **Figure 6.** Same as Fig. 4 but for the S cost function parameters (Table 6). Interannual variations
636 by the S model were well simulated for most sites, but less successfully for US-UMB and quite
637 poorly for US-Syv.

638

639 **Figure 7.** Linear regression derived slope of the relationship between annual NEE and anomaly
640 in dates of leaf on (L_{ON}), leaf off (L_{OFF}) and growing season length (G_{SL}) as quantified from
641 IFUSE model output using A_I cost function parameters (Table 5) plotted against linear
642 correlation of this relationship at all sites. Dotted line indicates $p < 0.1$ significance level.

643 **Tables**

644 **Table 1.** Location and types of co-located eddy covariance flux towers and time periods analyzed in this study along with their mean
 645 annual NEE and uncertainty in $\text{gC m}^{-2} \text{yr}^{-1}$, $1-\sigma$ IAV in $\text{gC m}^{-2} \text{yr}^{-1}$, post-filtering total gap fraction, and reference. Data gaps were
 646 strongly skewed toward nighttime data due to low turbulence screening criteria.

Name	Location	Type	Years	NEE	IAV	% Gaps	Reference
US-WCr	45° 48' N 90° 5' W	Forest	2000-2006	-380 ± 29	149	51%	Cook <i>et al.</i> , 2004
US-UMB	45° 34' N 84° 43' W	Forest	1999-2003	-284 +/- 22	43	39%	Gough <i>et al.</i> , 2008
US-Syv	46° 14' N 89° 21' W	Forest	2002-2006	1 ± 18	112	48%	Desai <i>et al.</i> , 2005
US-Los	46° 5' N 89° 59' W	Wetland	2001-2006	-84 ± 10	16	32%	Sulman <i>et al.</i> , 2009
US-PFa	45° 57' N 90° 16' W	Mixed	1997-2005	111 ± 21	67	28%	Ricciuto <i>et al.</i> , 2008

647

648 **Table 2.** Model parameters definitions, prior values, and acceptable posterior parameter ranges used by the IFUSE model and MCMC
 649 parameterization.

Name	Definition	Value
<i>Fixed parameters</i>		
k	Light extinction coefficient	0.5 FIXED
LA _{min}	Minimum leaf area	US-WCr 0.0, US-UMB 0.0, US-Syl 0.5, US-Los 0.0, US-PFa 0.5
LA _{max}	Maximum leaf area	US-WCr 5.3, US-UMB 3.7, US-Syl 4.1, US-Los 4.9, US-PFa 3.7
<i>Phenology parameters</i>		
α	Leaf on (L _{ON}) slope	0.05 (0.05-0.5)
GDD _{thresh}	Growing degree day threshold	200 (10-400)
β	Leaf off (L _{OFF}) slope	0.1 (0.05-0.5)
TEMP _{thresh}	Soil temperature threshold	4 (0-20)
<i>Photosynthesis parameters</i>		
LUE	Light use efficiency	0.25 (0-1)
T _{min}	Minimum photosynthetic temperature	0 (-15-10)
T _{opt}	Optimum photosynthetic temperature	15 (5-40)
VPD _{max}	Maximum photosynthetic VPD	3000 (0-20000)
VPD _{min}	Minimum photosynthetic VPD	100 (0-2000)
<i>Respiration parameters</i>		
r _s	Basal maintenance respiration	2 (0.1-5)
r _v	Basal growth respiration	2 (0.1-5)
b ₁	Maintenance respiration rate	0.03 (0-0.5)
b ₂	Growth respiration rate	0.03 (0-0.25)
b ₃	Leaf respiration fraction	0.05 (0-0.25)

650
 651

651 **Table 3.** Percent of variance explained (r^2) and mean absolute error (MAE) of observed NEE anomalies against the model in
652 asynchronous half-daily cost function (A_H), asynchronous interannual cost function (A_I), and synchronous (S) experiments. All
653 experiments were able to significantly capture daily variations NEE ($p < 0.01$), but A_I captured interannual variability at the most
654 number of sites, followed by the S. All correlations at the half-daily scale were significant, while significance of model-data
655 correlations of interannual variability are marked by *** ($p < 0.01$), ** ($p < 0.05$), and * ($p < 0.1$).

	Model	US-WCr	US-UMB	US-Syv	US-Los	US-PFa
Half-daily	r^2					
	A_H	0.90	0.93	0.87	0.88	0.83
	A_I	0.81	0.82	0.84	0.85	0.77
	S	0.81	0.88	0.73	0.82	0.77
	MAE					
	A_H	0.48	0.43	0.50	0.31	0.46
Inter-Annual	r^2					
	A_H	0.17	0.44	0.00	0.53	0.38
	A_I	0.87***	0.87***	0.69	0.76*	0.78**
	S	0.89***	0.64	0.24	0.88**	0.77**
	MAE					
	A_H	109	28	95	12	39
A_I	58	11	48	6	26	
S	46	31	71	9	25	

656

657

Table 4. Best and range of accepted posterior model parameters for each site using the asynchronous half-daily cost function (A_H).

Name	US-WCr	US-UMB	US-Syv	US-LoS	US-PFa
<i>Phenology parameters</i>					
α	0.089 (0.085-0.094)	0.142 (0.122-0.156)	0.195 (0.154-0.283)	0.206 (0.186-0.219)	0.160 (0.130-0.185)
GDD_{thresh}	154.0 (153.8-156.1)	107.1 (102.1-111.3)	81.1 (77.3-84.9)	80.7 (79.9-80.8)	131.5 (131.2-131.6)
β	0.182 (0.155-0.200)	0.165 (0.138-0.198)	0.098 (0.074-0.147)	0.091 (0.070-0.108)	0.135 (0.082-0.199)
$TEMP_{\text{thresh}}$	11.3 (11.1-11.8)	9.0 (8.9-9.2)	9.6 (9.2-11.1)	11.4 (11.3-11.4)	10.2 (9.2-10.9)
<i>Photosynthesis parameters</i>					
LUE	0.273 (0.262-0.281)	0.441 (0.326-0.444)	0.286 (0.265-0.371)	0.195 (0.187-0.251)	0.193 (0.179-0.200)
T_{min}	-5.6 (-15.0-1.7)	-15.0 (-15.0--14.2)	-14.0 (-15.0--10.7)	-11.0 (-14.9--7.1)	-3.5 (-7.7-0.4)
T_{opt}	6.7 (5.0-9.2)	39.1 (27.1-40.0)	29.5 (27.2-40.0)	32.1 (29.7-40.0)	11.0 (7.9-12.8)
VPD_{max}	4125 (3797-4562)	3143 (3016-3292)	3565 (3291-3908)	3893 (3636-4344)	3571 (3223-3929)
VPD_{min}	2 (0-115)	73 (0-309)	18 (0-235)	421 (15-513)	0 (0-201)
<i>Respiration parameters</i>					
r_s	1.01 (0.76-1.17)	0.11 (0.10-0.25)	0.12 (0.10-0.37)	0.55 (0.40-0.70)	0.73 (0.51-0.99)
r_v	0.22 (0.11-0.45)	0.95 (0.83-1.00)	1.55 (1.32-1.68)	0.65 (0.53-0.74)	0.71 (0.54-0.90)
b_1	0.0954 (0.0763-0.1238)	0.3812 (0.2634-0.4216)	0.1595 (0.0204-0.3538)	0.1091 (0.0923-0.1343)	0.1111 (0.0840-0.1405)
b_2	0.0275 (0.0002-0.0806)	0.0438 (0.0349-0.0499)	0.0811 (0.0730-0.0927)	0.0879 (0.0757-0.0983)	0.0872 (0.0723-0.1066)
b_3	0.0001 (0.0000-0.0052)	0.0007 (0.0000-0.0112)	0.0001 (0.0000-0.0391)	0.0121 (0.0001-0.0299)	0.0995 (0.0655-0.1221)

Table 5. Same as Table 4 but for the asynchronous interannual cost function (A_I).

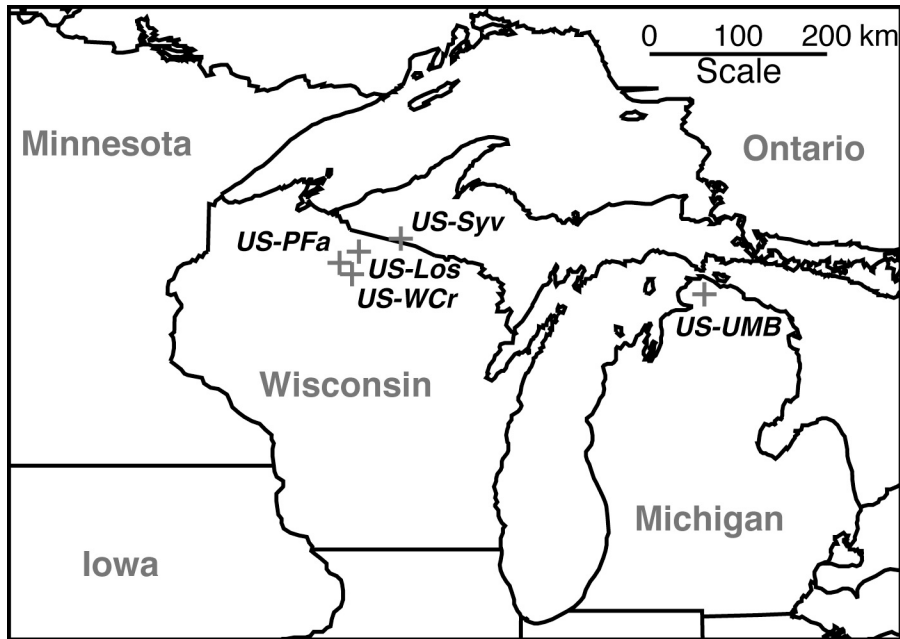
Name	US-WCr	US-UMB	US-Syv	US-LoS	US-PFa
<i>Phenology parameters</i>					
α	0.093 (0.084-0.097)	0.500 (0.467-0.500)	0.130 (0.113-0.139)	0.249 (0.245-0.251)	0.050 (0.050-0.051)
GDD _{thresh}	130.2 (127.6-130.3)	71.5 (71.4-71.5)	56.4 (56.3-56.9)	58.2 (57.1-58.4)	149.6 (149.4-149.8)
β	0.490 (0.317-0.500)	0.050 (0.050-0.053)	0.051 (0.050-0.059)	0.476 (0.450-0.483)	0.050 (0.050-0.050)
TEMP _{thresh}	1.4 (1.4-1.4)	-7.4 (-8.0--6.8)	14.6 (14.5-14.7)	8.3 (8.1-8.5)	-5.6 (-6.9--5.1)
<i>Photosynthesis parameters</i>					
LUE	0.463 (0.444-0.483)	0.204 (0.202-0.205)	0.244 (0.231-0.250)	0.183 (0.179-0.188)	0.159 (0.156-0.162)
T _{min}	-0.2 (-0.3--0.2)	8.9 (8.8-9.0)	6.7 (6.7-6.8)	-4.1 (-4.1--4.0)	5.0 (5.0-5.0)
T _{opt}	36.3 (34.6-37.8)	19.6 (19.5-19.7)	7.0 (6.9-7.0)	30.7 (30.3-31.5)	5.1 (5.1-5.2)
VPD _{max}	19802 (15935-19998)	2230 (2184-2301)	3681 (3489-3930)	4074 (3874-4314)	13304 (2110-19973)
VPD _{min}	10 (0-239)	1536 (1506-1549)	3 (0-209)	542 (480-603)	1964 (1732-1999)
<i>Respiration parameters</i>					
r _s	0.23 (0.22-0.24)	0.81 (0.78-0.81)	1.29 (1.21-1.41)	0.31 (0.28-0.33)	0.59 (0.55-0.63)
r _v	1.06 (1.04-1.12)	0.10 (0.10-0.12)	0.97 (0.86-1.05)	0.89 (0.86-0.92)	1.43 (1.39-1.47)
b ₁	0.4999 (0.4979-0.5000)	0.0449 (0.0414-0.0449)	0.1395 (0.1330-0.1461)	0.1859 (0.1729-0.2043)	0.0897 (0.0824-0.0944)
b ₂	0.0001 (0.0000-0.0011)	0.0002 (0.0000-0.0184)	0.0001 (0.0000-0.0020)	0.0769 (0.0745-0.0797)	0.0836 (0.0798-0.0883)
b ₃	0.0000 (0.0000-0.0010)	0.0005 (0.0000-0.0044)	0.0002 (0.0000-0.0066)	0.0004 (0.0000-0.0058)	0.0000 (0.0000-0.0016)

Table 6. Same as Table 4, but for the synchronous cost function (S).

Name	US-WCr	US-UMB	US-Syv	US-LoS	US-PFa
<i>Phenology parameters (jointly-optimized)</i>					
α	0.063 (0.061-0.065)				
GDD _{thresh}	165.3 (165.2-165.3)				
β	0.100 (0.084-0.112)				
TEMP _{thresh}	4.8 (4.7-4.8)				
<i>Photosynthesis parameters</i>					
LUE	0.299 (0.295-0.305)	0.495 (0.479-0.520)	0.300 (0.289-0.310)	0.260 (0.251-0.276)	0.159 (0.154-0.162)
T _{min}	4.0 (4.0-4.0)	-2.8 (-3.0--2.0)	9.4 (9.4-9.5)	2.7 (2.7-3.0)	-4.0 (-4.7--3.9)
T _{opt}	21.3 (21.0-21.6)	38.1 (36.9-39.9)	21.2 (21.0-21.8)	37.8 (36.4-40.0)	20.5 (19.4-21.2)
VPD _{max}	11931 (2010-19999)	3399 (3279-3647)	3389 (3261-3530)	4028 (3339-4565)	11206 (2442-19991)
VPD _{min}	1981 (1647-1999)	287 (99-369)	144 (10-229)	760 (721-956)	1943 (1084-1999)
<i>Respiration parameters</i>					
r _s	0.23 (0.22-0.24)	0.90 (0.81-0.92)	1.79 (1.74-1.81)	0.31 (0.28-0.52)	0.35 (0.26-0.45)
r _v	1.26 (1.23-1.29)	0.34 (0.30-0.39)	0.10 (0.10-0.12)	0.91 (0.71-0.93)	1.31 (1.26-1.37)
b ₁	0.4999 (0.4984-0.5000)	0.1073 (0.0987-0.1165)	0.2103 (0.2032-0.2146)	0.2720 (0.1780-0.3230)	0.2689 (0.2058-0.3217)
b ₂	0.0000 (0.0000-0.0009)	0.0002 (0.0000-0.0058)	0.0014 (0.0000-0.0187)	0.0859 (0.0635-0.0885)	0.0932 (0.0867-0.0991)
b ₃	0.0000 (0.0000-0.0005)	0.0003 (0.0000-0.0050)	0.0000 (0.0000-0.0035)	0.0004 (0.0000-0.0245)	0.0014 (0.0000-0.0142)

663 **Figures**

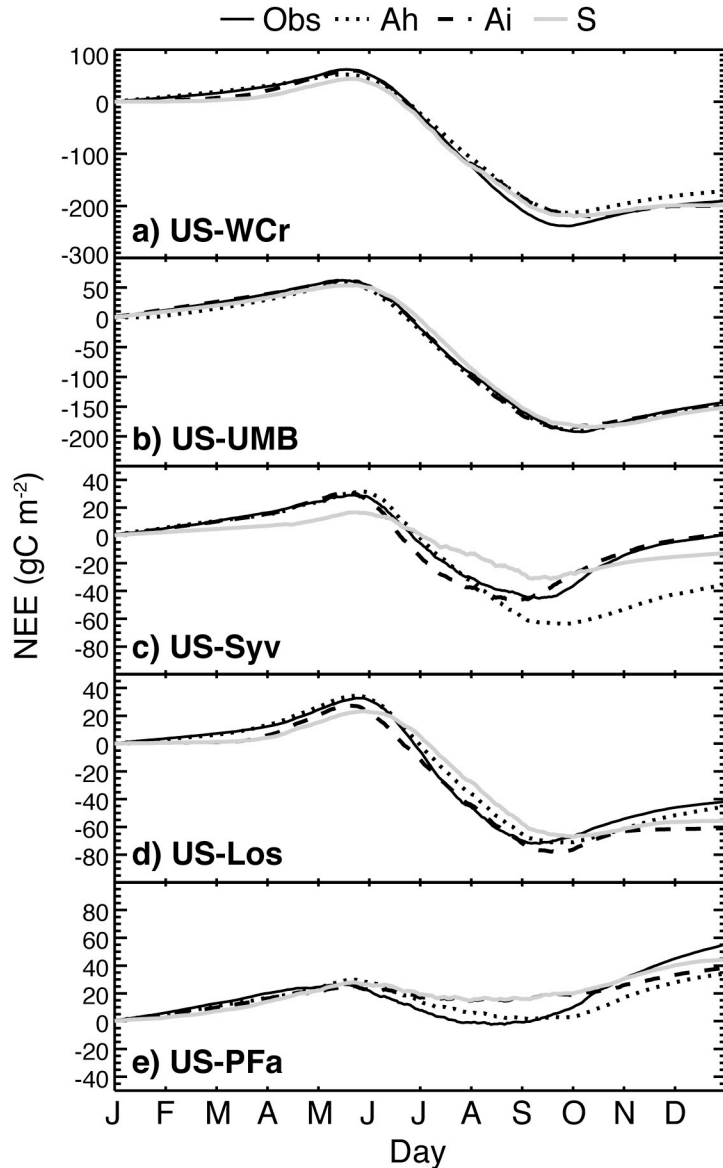
664 **Figure 1.** Map of north-central United States showing location of the five flux tower sites used
665 in this study.



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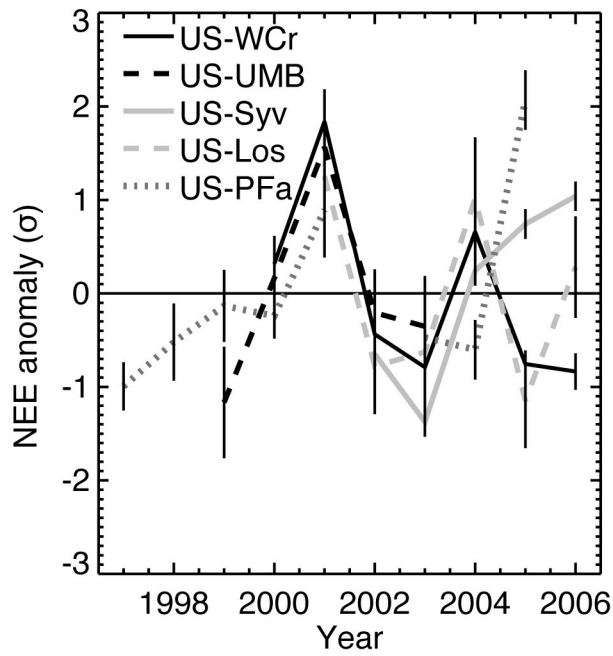
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667 **Figure 2.** Ensemble average cumulative daily NEE for observations, and the three model
 668 experiments (A_H , A_I , and S). Averages were performed over the observed record noted in Table
 669 1. The model generally captured the seasonal pattern of NEE at all sites, with worst performance
 670 at US-Syv and US-PFa. Observed and model uncertainty is not shown to preserve clarity, but
 671 generally fall within 10% of any observation.



672

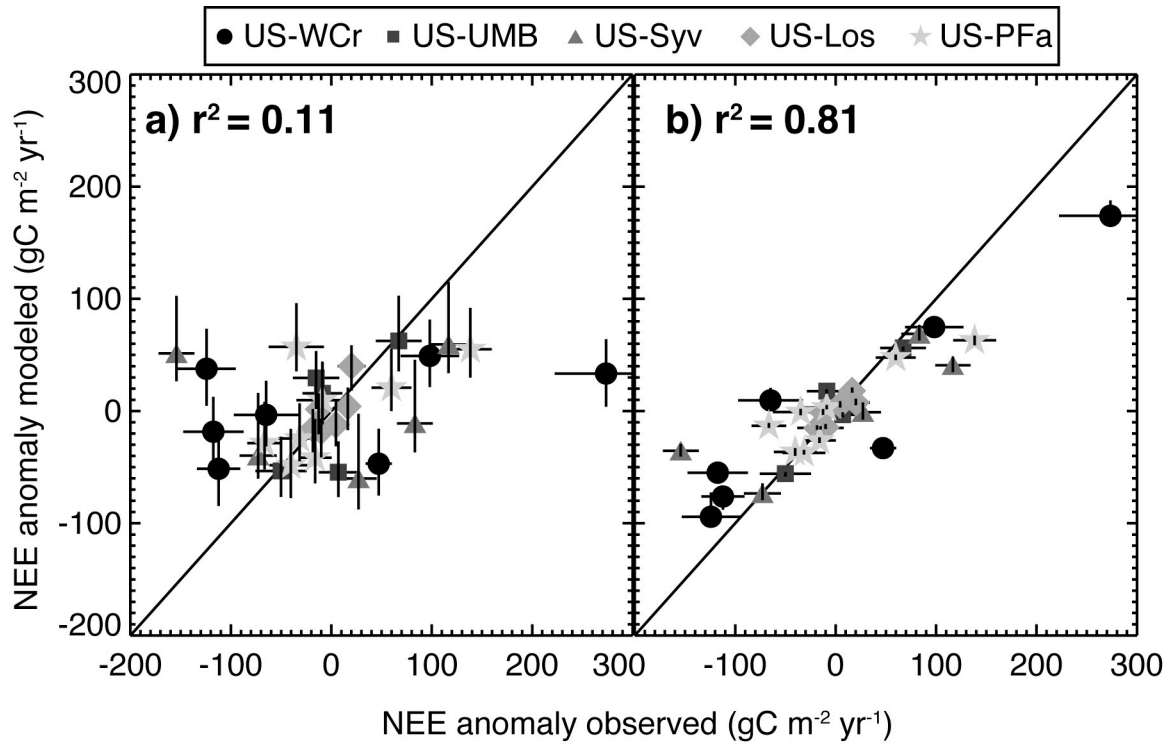
673 **Figure 3.** Observed standardized interannual variability in NEE at the five study sites. Strong
674 coherence in variability in NEE was observed across the time period, even though absolute
675 magnitudes in NEE variability varied widely. Observational uncertainty in NEE is noted by the
676 horizontal bars.



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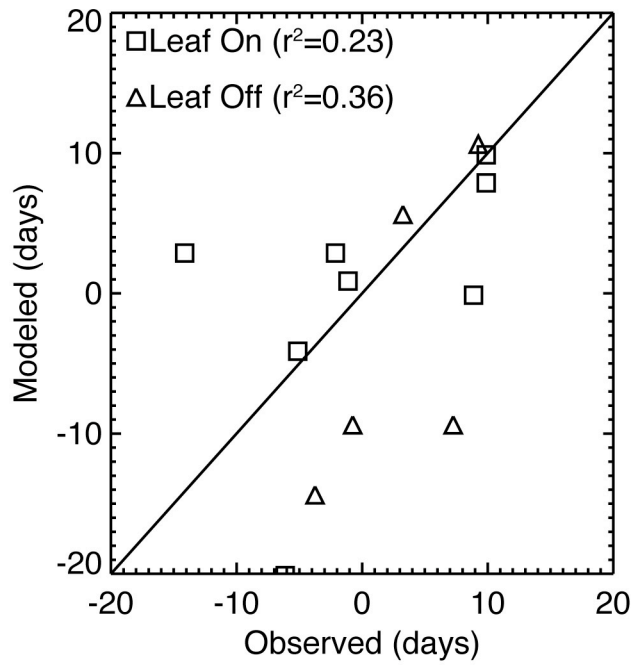
678 **Figure 4.** Correlation of anomalies in observed and modeled annual NEE using a) the A_H cost
679 function parameters (Table 4) and b) the A_I cost function parameters (Table 5). Significant
680 improvement in simulation of interannual variability was found for all sites in the latter.



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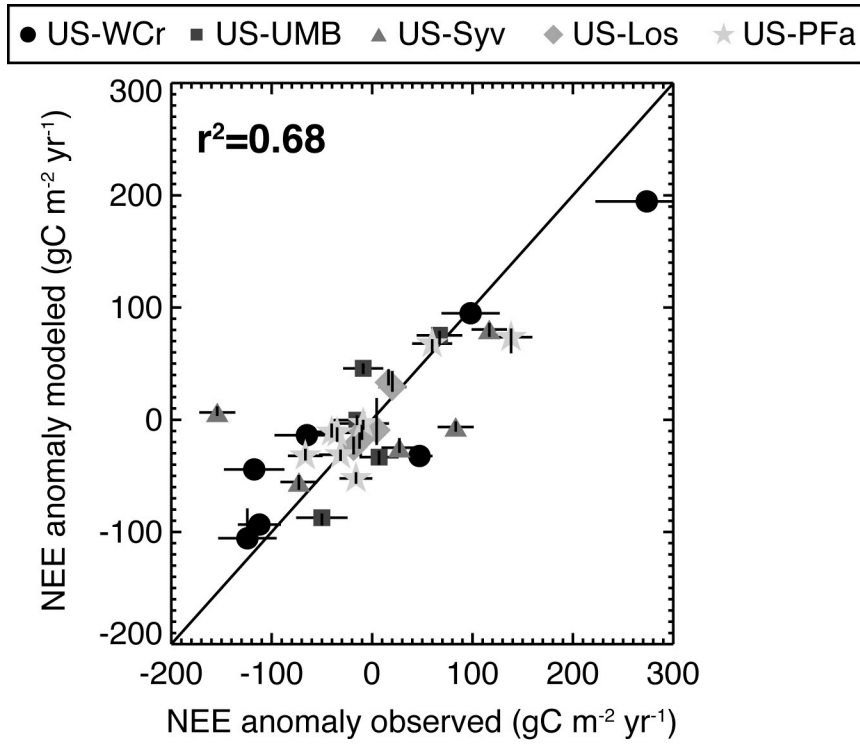
682 **Figure 5.** Comparison of observed light extinction profile derived leaf on (square) and leaf off
683 (triangle) date anomalies to IFUSE model for the US-WCr site, using the A_I parameters.
684 Generally, variability in both dates was modestly well simulated, though the slope of leaf off
685 appears too steep, while the leaf on dates mostly fall on the 1:1 line except for one outlier.



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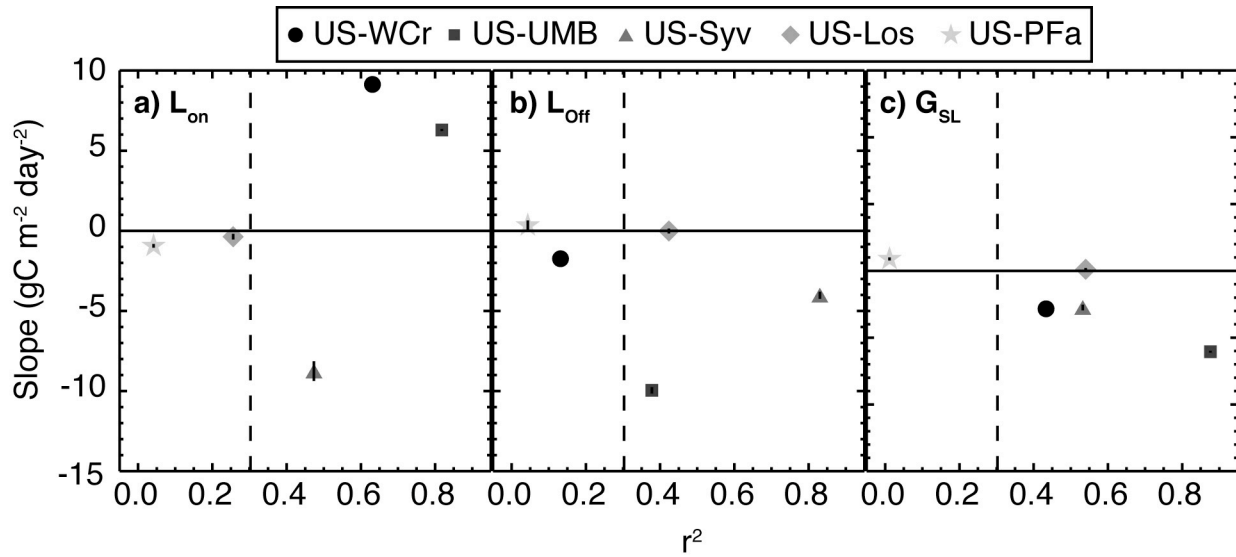
687 **Figure 6.** Same as Fig. 4 but for the S cost function parameters (Table 6). Interannual variations
688 by the S model were well simulated for most sites, but less successfully for US-UMB and quite
689 poorly for US-Syv.



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691

691 **Figure 7.** Linear regression derived slope of the relationship between annual NEE and anomaly
 692 in dates of leaf on (L_{ON}), leaf off (L_{OFF}) and growing season length (G_{SL}) as quantified from
 693 IFUSE model output using A_I cost function parameters (Table 5) plotted against linear
 694 correlation of this relationship at all sites. Dotted line indicates $p < 0.1$ significance level.



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